Experimental Determination of Stability Constant of Ferrous Iron Borate Complex [FeB(OH)₄⁺] at 25°C from Solubility Measurements

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10	
11	ABSTRACT
12	
13	In this study, the formation constant $(\log_{10} \beta_1^0)$ for FeB(OH) ₄ ⁺ as expressed in the
14	following reaction,
15	
16	$Fe^{2+} + B(OH)_4^- \Rightarrow FeB(OH)_4^+$
17	
18	is determined as 3.70 ± 0.10 (2 σ) at 25°C based on our long-term solubility experiments
19	up to 2,121 days on $Fe(OH)_2(cr)$ in the presence of borate ranging from 0.01 to 0.03
20	mol•kg ⁻¹ . In our solubility measurements, the equilibrium was attained from the direction of
21	supersaturation. In the experimental design, we used Fe ₂ (OH) ₃ Cl (pure iron end member of
22	hibbingite) as the starting material. When Fe ₂ (OH) ₃ Cl was in contact with solutions without
23	background concentrations of chloride, it was completely converted to Fe(OH) ₂ (cr).
24	The stability constant of $FeB(OH)_4^+$ is expected to find applications in many areas
25	of study. For instance, $FeB(OH)_4^+$ may have played an important role in transport of
26	ferrous iron in reducing water bodies at the surface of the primitive Earth. In the near-
27	field of geological repositories, the formation of $FeB(OH)_4^+$ can sequestrate soluble
28	borate, lowering borate concentrations available to the formation of the Am(III)-borate
29	aqueous complex.

32 1. INTRODUCTION

The formation constant for the ferrous iron borate complex, $FeB(OH)_{4^{+}}$, is not known (Bassett, 1980). In a critical review by Bassett (1980), there is a formation constant for the ferric iron borate complex, $FeB(OH)_{4^{2^{+}}}$, where the oxidation state for iron is +III, but there is no formation constant for $FeB(OH)_{4^{+}}$. The species $FeB(OH)_{4^{+}}$ is expected to be important in reducing environments. Hence, the presence of $FeB(OH)_{4^{+}}$ has relevance for many areas of study.

For instance in the early history of the Earth, before 2.4-2.3 billion years ago (2.4-2.3 Ga), oxygen was poor in the atmosphere (Holland, 1999), and therefore the ocean was anoxic, and iron was present as ferrous iron, i.e., Fe²⁺ (Pufahl and Hiatt, 2012). Based on boron isotope data and geological context, it has been inferred that ancient water bodies, such as oceans, were borate-rich, allowing evaporitic borate precursors to form before 2.1 Ga (e.g., Grew et al., 2011). The aqueous iron-borate complex, FeB(OH)₄⁺, could have played an important role in transporting iron in ancient, anoxic water bodies.

46 Another example is in the field of nuclear waste management. Waste containers 47 made of low-carbon steel, and in contact with water, or brine, are expected to corrode, 48 and may develop anoxic conditions, releasing ferrous iron in the near-field of a 49 geological repository. If the brine is in contact with the steel contains borate, ferrous iron 50 in solution can form $FeB(OH)_4^+$. The borate may originate either from the geological 51 formation or from corrosion of borosilicate glass waste forms (possibly used to dispose of 52 high level nuclear waste [HLW]). This iron aqueous complex could be a strong complex 53 and therefore compete with the formation of other borate complexes, sequestering free 54 borate in the system. A good example application is described below for the Waste

55	Isolation Pilot Plant (WIPP), a U.S. DOE geological repository for defense-related
56	transuranic (TRU) waste in the bedded salt formations in New Mexico, USA.
57	Borate can form aqueous complexes with Nd(III) and Eu(III) (Borkowski et al.,
58	2010; Schott et al., 2014, 2015), analogs for actinides in the +III oxidation state, such as
59	Am(III). If borate were to form an aqueous complex with Am(III) in the WIPP
60	repository, complexation with borate would contribute to the mobility of Am(III).
61	Complexation of Am(III) with borate would also increase the solubility of Am(III). The
62	inventory of Am(III) in waste was 143 kg for the WIPP Compliance Application Re-Certification
63	in 2009 (CRA-2009). The borate concentrations in the two WIPP brines that are important for
64	assessing performance of the repository, i.e., Generic Weep Brine (GWB), and Energy
65	Research and Development Administration (WIPP Well) 6 (ERDA-6), are 0.180 mol•kg ⁻¹ and
66	0.0692 mol•kg ⁻¹ (Xiong and Lord, 2008), respectively. Thus, the potential for formation of an
67	Am(III)/borate complex exists and sequestration of free borate in the $FeB(OH)_4^+$ complex
68	in WIPP should be evaluated. The stability constant for the $FeB(OH)_4^+$ complex is
69	needed for the evaluation.

The objective of this study is to determine the stability constant for $FeB(OH)_{4}^{+}$ via solubility measurements. The goal is two-fold: (1) to provide a quantitative tool to assess the role of $FeB(OH)_{4}^{+}$ in transportation of iron in the surface environments of the primitive Earth, and (2) to determine the stability constant for $FeB(OH)_{4}^{+}$ to use in assessment of performance of geological repositories .

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76 2. EXPERIMENTAL METHODS

In our experimental design, we used Fe₂(OH)₃Cl as the starting material. As was
 previously observed in our laboratory, if Fe₂(OH)₃Cl is added to a solution with zero

79	background Cl- concentration, then solid phase Fe(OH) ₂ forms. We considered that this
80	design, using the phase transformation, approaches the equilibrium from the direction of
81	supersaturation. The supporting solutions used were 0.01, 0.02, and 0.03 mol \cdot kg ⁻¹ H ₃ BO ₃ ,
82	as polyborate ions may form when total boron concentrations are higher than 0.03
83	mol•kg ⁻¹ (Mesmer et al., 1972). If polyborate ions are present, they will complicate data
84	interpretations. To prevent the problem of polyborate ions from our experiments, the
85	borate concentrations did not exceed 0.03 mol•kg ⁻¹ in our experimental design.
86	All experiments, including synthesizing the starting material and preparing
87	supporting solutions, were conducted in a VAC® glovebox (Model: Omni-Lab) with an
88	anoxic control system. Anoxic conditions were maintained using a source gas of 5 $\%~H_2$
89	(balance Ar) and O ₂ scrubber boxes manufactured by VAC®. The oxygen concentration
90	within the glovebox was maintained below 1 ppm, routinely around 0.2 ppm. Prior to
91	preparing solutions, deionized water (DI) with 18.3 M Ω was sparged with the anoxic-
92	glovebox gas in the glovebox for a minimum of one hour.
93	
94	2.1 Synthesis of Starting Material
95	In our synthesis of the starting material, Fe ₂ (OH) ₃ Cl, we followed a method
96	similar to that used in our laboratory before, as described in Nemer et al. (2011). In the
97	synthesis, we first dissolved 48 grams of FeCl2•4H2O(s) (Fisher Scientific, ACS grade)
98	into 100 mL degassed DI water. Second, we dissolved 41 grams of NaOH(s) (Fisher
99	Scientific, ACS grade) into 60 mL degassed DI water. Third, 26 mL of the NaOH
100	solution described above was transferred into the FeCl ₂ solution described above by using
101	a graduated cylinder. Then, 26 mL of degassed water was used to rinse the graduated

102 cylinder used in the previous step. The rinsed solution was also transferred into the
103 above FeCl₂ solution. Finally, the plastic bottle containing the FeCl₂ solution was doubly
104 wrapped with Teflon tape, first on the thread, and then on the lid. We shook the bottle
105 well before storing it. After some time (in this case, about a year), the precipitate was
106 filtered out.

107 In the filtration process based on gravity, after the filter paper and filtering 108 apparatus was set-up, the slurry containing the precipitate was poured into the filtering 109 apparatus, and the bottle was rinsed thoroughly with DI water until the bottle was clear. 110 The rinsing DI water was also poured into the filtering apparatus. The solid was washed 111 several times. The slurry was left on the filtering apparatus for a few days to ensure that 112 the solid was completely filtered and to allow drying of the solid. In the filtration 113 process, we tried two types of filter papers. We first used #40 Whatman® filter paper. 114 However, the filtrate looked cloudy after filtration for two hours. Therefore, the filtrate 115 was collected and poured back into the filtering apparatus to make sure that no 116 Fe₂(OH)₃Cl was lost. Then we used Millipore-Isopore[®] membrane 1.2 µm filter paper. 117 This filter paper was better, as the filtrate was very clear.

118

119 2.2 Experimental Setup

Supersaturation experiments using the phase transformation from $Fe_2(OH)_3Cl(cr)$ to $Fe(OH)_2(cr)$, mentioned before, are conducted at 25.0 ± 0.5 C. In our solubility experiments, approximately 1.5 grams of $Fe_2(OH)_3Cl(cr)$ were placed into serum bottles along with 100 mL of supporting solutions with the desired borate concentrations. The supporting solutions consisted of 0.01, 0.02, and 0.03 mol•kg⁻¹ H₃BO₃. All supporting

solutions were prepared from ACS grade H₃BO₃ from Fisher Scientific, and deaerated DI
water.

127 Solution samples were periodically withdrawn from the experiments to determine 128 if the system had reached equilibrium. Before each sampling, pH readings were taken for 129 each experiment. In each sampling, about 3 mL of solution samples were taken from 130 each experiment, and the solution samples were filtered through a 0.2 µm filter, and 131 transferred into pre-weighed 10 mL Grade A volumetric flasks. After filtration, masses 132 of each solution sample were determined with a balance precise to the fourth decimal 133 place. Samples were then immediately acidified with 0.5 mL of the Optima® Grade 134 HNO₃ from Fisher Scientific, and diluted to 10 mL with DI water. Prior to chemical 135 analyses for iron and boron using the PerkinElmer Optima 3300 Dual View (DV) or 136 PerkinElmer Optima 8300 Dual View (DV) ICP-AES, aliquots from the afore-mentioned 137 acidified samples were further diluted to an appropriate ionic strength. 138 The pH readings were measured with an Orion-Ross combination pH glass 139 electrode, coupled with an Orion Research EA 940 pH meter that was calibrated with 140 three pH buffers (pH 4, pH 7, and pH 10). The pH scale used in this work is the 141 concentration scale (Mesmer and Holmes, 1992), denoted as pH_m. 142 Chloride concentrations were determined with a DIONEX ion chromatograph 143 (IC) (DIONEX IC 3000) with a conductivity detector. 144 The final measurements for this study included iron, chloride, boron, and 145 hydrogen ion molal concentration data.

146 Solid phases were analyzed using a Bruker D8 Advance X-ray diffractometer

147 with a Sol-X detector, before and after experiments. XRD patterns were collected using

148 CuK α radiation at a scanning rate of 1.33°/min for a 2 θ range of 10–90°.

149

150 3. RESULTS

Experimental results are tabulated in Table 1. In Figure 1, the XRD patterns are presented. In the lower portion of Figure 1, it is shown that the starting material is $Fe_2(OH)_3Cl$ with a minor amount of $Fe(OH)_2(cr)$. In the upper portion of Figure 1, the XRD patterns for the solids after experiment are displayed. These XRD patterns indicate the complete transformation from $Fe_2(OH)_3Cl$ to $Fe(OH)_2(cr)$, demonstrating that $Fe(OH)_2(cr)$ was the solubility-controlling phase.

157 In Figure 2, negative logarithms of hydrogen ion concentrations (pH_m) as a 158 function of experimental time are shown. For the purpose of comparison with a similar 159 system on brucite $[Mg(OH)_2(cr)]$, the evolution of pH_m in a long-term experiment 160 approaching equilibrium with brucite from supersaturation (Xiong, 2008) is also 161 displayed in Figure 2. The trends are similar for both systems. In the case of brucite, the 162 reversal, i.e., the results from the experiment approaching equilibrium from the direction 163 of supersaturation being the same with the results from the direction of undersaturation, 164 was attained at ~83 days (Xiong, 2008). The experiments from both directions were 165 conducted in that work. In the case of $Fe(OH)_2(cr)$, it seems that the steady state was 166 attained starting from, or after, the second sampling (i.e., 108 days, see Table 1), when 167 the evolution of iron concentrations and the solubility quotients as a function of 168 experimental time described below are taken into consideration.

- In Figure 3, molalities of dissolved ferrous iron as a function of experimental time are presented. It shows that the iron concentrations in the first sampling (i.e., 37 days) are much higher than those in the rest of the samplings, indicating that the system was still oversaturated with respect to Fe(OH)₂(cr). This is similar to the evolution of magnesium concentrations and solubility products.
- 174 In Figure 4, the solubility quotients in logarithmic units are plotted versus 175 experimental time. In the pH_m range of our experiments, Fe^{2+} is the dominant species. 176 The dissolution reaction for $Fe(OH)_2(cr)$ can be generally expressed as follows,
- 177

178
$$Fe(OH)_2(cr) + 2H^+ \Rightarrow Fe^{2+} + 2H_2O(l)$$
 (1)

180
$$Q = \frac{m_{\Sigma Fe(II)}}{(m_{H^+})^2}$$
(2)

181

Figure 4 shows that the steady state was attained starting from the second or third sampling. In our previous work on $Fe(OH)_2(cr)$ from the direction of undersaturation, the steady state was attained no later than 94 days (Nemer et al. 2011). Combining this work on $Fe(OH)_2(cr)$ from supersaturation in the presence of borate with our previous work on $Fe(OH)_2(cr)$ from undersaturation without borate, we infer that the reversal for the solubility of $Fe(OH)_2(cr)$ was attained in ~100-200 days. This time scale is similar to that for the solubility of brucite, which attained the reversal at ~83 days (Xiong, 2008).

190 4. THERMODYNAMIC MODEL, DISCUSSIONS, AND APPLICATIONS

192 4.1 Thermodynamic Modeling

In the hydrolysis speciation scheme for Fe(II) recommended by Baes and Mesmer (1976), Fe²⁺ is dominant at pH \leq 9, followed by FeOH⁺ at 9 < pH \leq 10.5, by Fe(OH)₃⁻ at pH > 10.5. Fe(OH)₂(aq) and Fe(OH)₄²⁻ are weak species, and they do not have their dominant fields (Baes and Mesmer, 1976). The formation of FeOH⁺ can be expressed as, 197

198
$$Fe^{2+} + H_2O(1) \Rightarrow FeOH^+ + H^+$$
 (3)

199

200
$$K_{1,1}^{0} = \frac{(m_{H^{+}}) \times (m_{FeOH^{+}})}{m_{Fe^{2^{+}}}} \times \frac{(\gamma_{H^{+}}) \times (\gamma_{FeOH^{+}})}{\gamma_{Fe^{2^{+}}}}$$
(4)

201

where m_i is a molal concentration for the *i*-th species, γ_i an activity coefficient for the *i*-th species.

In the presence of borate, $FeB(OH)_4^+$ can form according to the following reaction,

206

207
$$\operatorname{Fe}^{2+} + \operatorname{B}(\operatorname{OH})_4^- \rightleftharpoons \operatorname{FeB}(\operatorname{OH})_4^+$$
 (5)

208

209
$$\beta_{1}^{0} = \frac{m_{FeB(OH)_{4}^{+}}}{(m_{Fe^{2^{+}}}) \times (m_{B(OH)_{4}^{-}})} \times \frac{\gamma_{FeB(OH)_{4}^{+}}}{(\gamma_{Fe^{2^{+}}}) \times (\gamma_{B(OH)_{4}^{-}})}$$
(6)

211 Therefore, the iron species considered in this study for modeling include Fe^{2+} ,

212 FeOH⁺ and FeB(OH)₄⁺, according to the experimental pH_m 's in this work.

213 Based on our experimental data, we model the equilibrium constant at infinite 214 dilution for Reaction (5). In the modeling, we use the computer code EQ3/6 Version 8.0a 215 (Wolery et al. 2010; Xiong 2011) as a modeling platform. The essence of the modeling is 216 to minimize the difference between experimental and model predicted values. The 217 EQ3/6 code has been successfully used as a modeling platform in a number of previous 218 studies at ambient temperature (e.g., Xu et al., 1999; Kong et al., 2013; Xiong et al. 219 2013a, 2013b, 2017; Xiong 2015) and at elevated temperatures up to 523.15 K (Xiong 220 2013a, 2013b, 2014).

221 Experimental data after attainment of the steady state are selected for the 222 modeling. Those data are used to generate EQ3/6 input files in the first step. In the second step, the values of the targeted parameter (e.g., $\log_{10} \beta_1^0$) are changed until the 223 difference between experimental values (e.g., $m_{\Sigma Fe(II)}$) and model-produced values 224 225 predicted by the computer code is minimized. In the third step, the final selected value 226 for the targeted parameter is verified to see how it reproduces experimental data. 227 The database used for the modeling is DATA0.FM2 (Xiong and Domski, 2016), 228 which utilizes the Pitzer model for calculations of activity coefficients of aqueous 229 species. The original borate chemistry in DATA0.FM2 was from Felmy and Weare 230 (1986), with updates from Xiong et al. (2013). The equilibrium constant for Reaction (1) 231 is from Nemer et al. (2011) and its recent revision (Domski, 2017), and the first 232 hydrolysis constant for Reaction (3) is from the EQ3/6 database, DATA0.YMP (Wolery 233 and Jarek, 2003).

234	Based on the modeling, the equilibrium constant for Reaction (5) is determined to
235	be 3.70 ± 0.10 (2 σ) in 10-based logarithmic units (Table 2). In Figure 5, the model-
236	predicted solubilities of $Fe(OH)_2(cr)$ in the presence of borate at 0.01, 0.02, and 0.03
237	mol \cdot kg ⁻¹ are compared with the experimental values. Two things are clear from Figure
238	5. First, solubilities of $Fe(OH)_2(cr)$ in the presence of borate are much higher than those
239	without borate, suggesting that borate forms an aqueous complex, $FeB(OH)_4^+$, with $Fe(II)$
240	to enhance solubilities of Fe(OH) ₂ (cr). Second, Figure 5 demonstrates that the values
241	predicted using the model with $FeB(OH)_4^+$ are in good agreement with the experimental
242	values.
243	Van den Berg (1984) determined the equilibrium constants for $CuB(OH)_4^+$ and
244	PbB(OH) ₄ ⁺ at 25°C and $I = 0.7$ mol•dm ⁻³ KNO ₃ as 3.48 and 2.2 in 10-based logarithmic
245	units, respectively. When these conditional equilibrium constants are extrapolated to
246	infinite dilution using the B dot equation (Helgeson, 1969), they are 4.19 and 2.91,
247	respectively. In the extrapolation to infinite dilution, the ionic strength on a molar scale
248	is converted to one (i.e., $0.72 \text{ mol} \cdot \text{kg}^{-1}$) on a molal scale by using the density equation for
249	KNO ₃ from Sőhnel and Novotný (1985). Therefore, the strength of $FeB(OH)_4^+$ is
250	between those for $CuB(OH)_4^+$ and $PbB(OH)_4^+$.
251	
252	4.2 Applications

Before the Great Oxidation Event (GOE) that happened before 2.4-2.3 Ga, the partial pressure of oxygen, P_{O_2} , was lower than 10^{-4} atm in the Earth's atmosphere (Bekker and Holland, 2012). Under such reducing conditions, dissolved iron was present as ferrous iron in the primitive oceans (Pufahl and Hiatt, 2012). Under the pH conditions

257	of the primitive oceans that was close to neutral, but slightly alkaline, $FeB(OH)_4^+$ may
258	have played an important role in transport of iron at the surface of the primitive Earth, as
259	borate is expected to be present in the primitive oceans at significant concentrations to
260	have formed evaporitic borate precursors (e.g., Grew et al., 2011).
261	The only competitor for $FeB(OH)_4^+$ in transport of iron in the primitive oceans
262	may have been $FeHCO_3^+$. The formation of $FeHCO_3^+$ can be expressed as follows,
263	
264	$Fe^{2+} + HCO_3^- \Rightarrow FeHCO_3^+$ (7)
265	
266	The equilibrium constant for Reaction (7) is 2.72 in 10-based logarithmic units (Mattigod
267	and Sposito, 1979). In comparison with $\log_{10} \beta_1^0$ of $3.70 \pm 0.10 \ (2\sigma)$ for FeB(OH) ₄ ⁺ , the
268	iron complex with bicarbonate is weaker. Therefore, $FeB(OH)_4^+$ may have played a
269	dominant role in transport of iron in the primitive oceans.
270	In geological repositories for disposal of nuclear waste, iron is present in
271	repositories as waste containers and waste, and borate concentrations from the geological
272	formations and degradation of borosilicate waste forms for high level nuclear waste
273	(HLW) can be significant. For instance, tincalconite (Na ₂ B ₄ O ₇ •5H ₂ O) has been observed as
274	one corrosion product for borosilicate glass for HLW under repository conditions in China
275	(Zhang et al., 2012). The formation of tincalconite suggests that the borate concentrations are
276	high enough to reach the saturation limit of tincalconite. As borate can form a complex with
277	Am(III), AmHB ₄ O ₇ ²⁺ (Borkowski et al., 2010; Xiong, 2017), the formation of such a Am(III)-
278	borate complex could contribute to the higher solubility of Am(III). However, under the reducing

borate complex could contribute to the higher solubility of Am(III). However, under the reducing

279 conditions of geological repositories, the formation of $FeB(OH)_4^+$ will alleviate the effect of 280 borate complexation with Am(III).

281

282 5. SUMMARY

In this work, we determine the stability constant of $FeB(OH)_4^+$ based on solubility measurements on $Fe(OH)_2(cr)$ in the presence of borate. The strength of this complex is between those for $CuB(OH)_4^+$ and $PbB(OH)_4^+$. It is expected that $FeB(OH)_4^+$ plays an important role under reducing environments for iron chemistry.

287

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421 Table 1. Experimental results produced in this study at $25.0 \pm 0.5^{\circ}$ C.

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	Supporting			Molality of total
	Medium,	Experimental		dissolved ferrous iron,
Experimental Number	H ₃ BO ₃ , molal	time, days	pН	$m_{\Sigma Fe(II)}$
FeB(OH)4-0.01B-1	0.01	37	7.88	1.24E-02
FeB(OH)4-0.01B-2	0.01	37	7.88	1.16E-02
FeB(OH)4-0.02B-1	0.02	37	7.94	8.49E-03
FeB(OH)4-0.02B-2	0.02	37	7.95	8.75E-03
FeB(OH)4-0.03B-1	0.03	37	8.02	7.60E-03
FeB(OH)4-0.03B-2	0.03	37	8.03	7.38E-03
FeB(OH)4-0.01B-1	0.01	108	7.97	4.01E-03
FeB(OH)4-0.01B-2	0.01	108	7.96	7.84E-03
FeB(OH)4-0.02B-1	0.02	108	8.00	7.30E-03
FeB(OH)4-0.02B-2	0.02	108	8.08	5.65E-03
FeB(OH)4-0.03B-1	0.03	108	8.11	6.10E-03
FeB(OH)4-0.03B-2	0.03	108	8.12	6.55E-03
FeB(OH)4-0.01B-1	0.01	184	7.83	5.20E-03
FeB(OH)4-0.01B-2	0.01	184	7.95	3.84E-03
FeB(OH)4-0.02B-1	0.02	184	7.95	4.78E-03
FeB(OH)4-0.02B-2	0.02	184	7.94	3.61E-03
FeB(OH)4-0.03B-1	0.03	184	7.97	4.44E-03
FeB(OH)4-0.03B-2	0.03	184	8.04	2.75E-03
FeB(OH)4-0.01B-1	0.01	457	7.87	6.59E-03
FeB(OH)4-0.01B-2	0.01	457	7.88	4.95E-03
FeB(OH)4-0.02B-1	0.02	457	7.94	6.25E-03
FeB(OH)4-0.02B-2	0.02	457	7.94	3.87E-03
FeB(OH)4-0.03B-1	0.03	457	7.96	6.09E-03
FeB(OH)4-0.03B-2	0.03	457	7.98	5.12E-03
FeB(OH)4-0.01B-1	0.01	496	7.84	6.82E-03
FeB(OH)4-0.01B-2	0.01	496	7.83	5.12E-03
FeB(OH)4-0.02B-1	0.02	496	7.85	6.55E-03
FeB(OH)4-0.02B-2	0.02	496	7.84	4.07E-03
FeB(OH)4-0.03B-1	0.03	496	7.67	6.41E-03
FeB(OH)4-0.03B-2	0.03	496	7.90	5.24E-03

FeB(OH)4-0.01B-1	0.01	576	7.85	6.85E-03
FeB(OH)4-0.01B-2	0.01	576	7.85	5.10E-03
FeB(OH)4-0.02B-1	0.02	576	7.88	6.71E-03
FeB(OH)4-0.02B-2	0.02	576	7.88	4.03E-03
FeB(OH)4-0.03B-1	0.03	576	7.88	6.40E-03
FeB(OH)4-0.03B-2	0.03	576	7.90	5.38E-03
FeB(OH)4-0.01B-1	0.01	616	7.77	6.83E-03
FeB(OH)4-0.01B-2	0.01	616	7.80	5.22E-03
FeB(OH)4-0.02B-1	0.02	616	7.83	6.62E-03
FeB(OH)4-0.02B-2	0.02	616	7.82	3.94E-03
FeB(OH)4-0.03B-1	0.03	616	7.85	6.26E-03
FeB(OH)4-0.03B-2	0.03	616	7.87	5.36E-03
FeB(OH)4-0.01B-1	0.01	1547	7.73	7.54E-03
FeB(OH)4-0.01B-2	0.01	1547	7.78	5.48E-03
FeB(OH)4-0.02B-1	0.02	1547	7.80	7.29E-03
FeB(OH)4-0.02B-2	0.02	1547	7.86	3.33E-03
FeB(OH)4-0.03B-1	0.03	1547	7.82	7.22E-03
FeB(OH)4-0.03B-2	0.03	1547	7.81	6.28E-03
FeB(OH)4-0.01B-1	0.01	1617	7.75	7.78E-03
FeB(OH)4-0.01B-2	0.01	1617	7.80	5.42E-03
FeB(OH)4-0.02B-1	0.02	1617	7.82	7.26E-03
FeB(OH)4-0.02B-2	0.02	1617	7.90	3.31E-03
FeB(OH)4-0.03B-1	0.03	1617	7.82	7.20E-03
FeB(OH)4-0.03B-2	0.03	1617	7.83	6.30E-03
FeB(OH)4-0.01B-1	0.01	1743	7.70	7.82E-03
FeB(OH)4-0.01B-2	0.01	1743	7.78	5.37E-03
FeB(OH)4-0.02B-1	0.02	1743	7.81	7.19E-03
FeB(OH)4-0.02B-2	0.02	1743	7.88	3.19E-03
FeB(OH)4-0.03B-1	0.03	1743	7.80	7.16E-03
FeB(OH)4-0.03B-2	0.03	1743	7.81	6.28E-03
FeB(OH)4-0.01B-1	0.01	1841	7.80	8.11E-03
FeB(OH)4-0.01B-2	0.01	1841	7.81	5.46E-03
FeB(OH)4-0.02B-1	0.02	1841	7.91	7.74E-03
FeB(OH)4-0.02B-2	0.02	1841	7.89	3.34E-03

0.03	1841	7.85	7.69E-03
0.03	1841	7.90	6.78E-03
0.01	1868	7.76	8.30E-03
0.01	1868	7.81	5.43E-03
0.02	1868	7.84	7.60E-03
0.02	1868	7.92	3.20E-03
0.03	1868	7.81	7.57E-03
0.03	1868	7.82	6.63E-03
0.01	2024	7.69	8.29E-03
0.01	2024	7.77	5.26E-03
0.02	2024	7.76	7.79E-03
0.02	2024	7.89	3.11E-03
0.03	2024	7.77	7.74E-03
0.03	2024	7.78	6.74E-03
0.01	2121	7.70	8.15E-03
0.01	2121	7.78	4.96E-03
0.02	2121	7.76	7.25E-03
0.02	2121	7.86	2.95E-03
0.03	2121	7.75	7.54E-03
0.03	2121	7.74	6.83E-03
	0.03 0.03 0.01 0.01 0.02 0.02 0.03 0.01 0.01 0.02 0.02 0.03 0.03 0.01 0.01 0.02 0.03 0.01 0.02 0.03 0.01 0.03 0.03	$\begin{array}{cccccccccccccccccccccccccccccccccccc$	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$

426 427 Table 2. The equilibrium constants for the system Fe^{2+} —OH⁻—B(OH)₄⁻ at 25°C

Reactions	$\log_{10} \mathrm{K}^{0}$	Reference
$Fe(OH)_2(cr) + 2H^+ \Rightarrow Fe^{2+} + 2H_2O(1)$		Nemer et al. (2011),
	$12.89 \pm 0.13 \ (2\sigma)$	Domski (2017)
$Fe^{2+} + H_2O(1) \Rightarrow FeOH^+ + H^+$	-9.3148	DATA0.YMP*
$Fe^{2+} + B(OH)_4^- \Rightarrow FeB(OH)_4^+$	$3.70 \pm 0.10 (2\sigma)$	This Study
* Wolery and Jarek (2003)		

Wolery and Jarek (2003)

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432	Figure Captions
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435	Figure 1. The XRD pattern for a representative sample after experiments is displayed.
436	An XRD pattern for the synthetic Fe(OH) ₂ (cr) produced in this work with the reference
437	standard of Fe(OH) ₂ (cr) from PDF-4+ 2016 (Software Version 4.16.04, Database Version
438	4.1605) of the International Center for Diffraction Data (ICDD) is also presented.
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441	Figure 2. A plot showing hydrogen ion concentrations on a molal scale (pH _m) measured
442	by this study as a function of experimental time.
443	
444	
445	Figure 3. A plot showing Fe(II) concentrations as a function of experimental time.
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447	
448	Figure 4. A plot showing $log Q$ as a function of experimental time
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450	
451	Figure 5. Comparisons of experimental solubilities of Fe(OH) ₂ (cr) in the presence of
452	borate with the solubilities of $Fe(OH)_2(cr)$ predicted by using models with $FeB(OH)_4^+$
453	and without $FeB(OH)_4^+$, as a function of hydrogen ion concentrations on a molal scale.
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463 Figure 1.







